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Abstract

The Authie estuarie, located at the eastern part of the English Channel is of environmental, ecological, economic and societal importance. With the intention to better understand the sediment dynamic it is important to better assess the role of sediment dynamics including erosion, stabilization and sediment reworking processes which is challenging in such complex environment. It is also important to consider biogenic components such as the microphytobenthos (MPB) distribution, as the primary productivity may play an important role with the bio-stabilization process. As a consequence, there is a crucial need to provide a synoptic overview of inherent bio-physical characteristics of sediments (i.e., composition, water content, grain-size, and biomass) in estuarine environment by generating precise quantitative maps for predicting in a second step estuarine evolution by including sediment transport, sedimentation rates, coastal flows processes and sea level rise caused by climate change for instance. The use of the remote sensing technology is increasingly used for mapping estuarine and coastal environments by providing a synoptic overview of bio-physical characteristics of sediments. In that sense, the combination between remote sensing imaging, topographic data (LiDAR) and *in situ* measurements is suitable for improving our understanding of sediment dynamics with respect to physical and biological forcings. The main objective of this study is to demonstrate that the synergy between multispectral (i.e., SPOT 6–7 [1.5 m/pixel]; Sentinel-2, 10–60 m/pixel, 5–10 days)”, hyperspectral [Hypex, 70 cm/pixel, 160 spectral bands] remote sensing images may be suitable for generating both reliable sedimentary and primary productivity budgets; at least for surficial sediments. All presented data were acquired during the same day (09/21/2017) in the framework TéléEST, CPER MARCO and CNRS-OMPBI projects.

Keywords
(separated by '-')

Remote sensing - Hyperspectral - Multispectral - LiDAR - Physical properties mapping - Morphology - Bay of authie

Synergy Between Hyperspectral (HYSPEX), Multispectral (SPOT 6/7, Sentinel-2) Remotely Sensed Data and LiDAR Data for Mapping the Authie Estuarie (France)



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Abstract The Authie estuarie, located at the eastern part of the English Channel is of environmental, ecological, economic and societal importance. With the intention to better understand the sediment dynamic it is important to better assess the role of sediment dynamics including erosion, stabilization and sediment reworking processes which is challenging in such complex environment. It is also important to consider biogenic components such as the microphytobenthos (MPB) distribution, as the primary productivity may play an important role with the bio-stabilization process. As a consequence, there is a crucial need to provide a synoptic overview of inherent bio-physical characteristics of sediments (i.e., composition, water content, grain-size, and biomass) in estuarine environment by generating precise quantitative maps for predicting in a second step estuarine evolution by including sediment transport, sedimentation rates, coastal flows processes and sea level rise caused by climate change for instance. The use of the remote sensing technology is increasingly used for mapping estuarine and coastal environments by providing a synoptic overview of bio-physical characteristics of sediments. In that sense, the combination between remote sensing imaging, topographic data (LiDAR) and *in situ* measurements is suitable for improving our understanding of sediment dynamics with respect to physical and biological forcings. The main objective of this study is to demonstrate

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19 that the synergy between multispectral (i.e., SPOT 6–7 [1.5 m/pixel]; Sentinel-2, 10–
20 60 m/pixel, 5–10 days)”, hyperspectral [Hypex, 70 cm/pixel, 160 spectral bands]
21 remote sensing images may be suitable for generating both reliable sedimentary
22 and primary productivity budgets; at least for surficial sediments. All presented data
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24 MARCO and CNRS-OMPBI projects.

25 **Keywords** Remote sensing · Hyperspectral · Multispectral · LiDAR · Physical
26 properties mapping · Morphology · Bay of Authie

27 1 Introduction

28 With the intention to improve our understanding of sediment dynamics within estu-
29arine environments such as the Bay of Authie (North of France) it is crucial to
30 generate precise quantitative maps of physical and biological properties of the sedi-
31ment deposits. Inherent sedimentary parameters such as grain-size, moisture content,
32 biological content, topography as well as physical forcing conditions (e.g. tidal con-
33 ditions and wave currents) are needed for retrieving environmental changes over
34 time and therefore to be able to generate reliable sedimentary budgets prior to define
35 any management strategies or any decisions which may help to overcome erosion
36 or sandfilling issues. Sedimentary trend predictions are generally based on field data
37 such as samples sediments, in situ measurements but also any precise quantitative
38 maps. In that sense remotely sensed products, maps of sedimentary parameters, may
39 be assimilated by models such as transport models, erosion/stabilization models and
40 thus for computing those predictions under various scenario. Even hydrodynamic
41 condition knowledge is needed for fully assessing the general sedimentary trend, in
42 this study, we decided to only focus on inherent sedimentary parameters mapping.
43 Grain-size maps are important for fully assessing sediment transport trend, erodibil-
44 ity of sediment surfaces, etc. It is also important to consider moisture content both in
45 sedimentology aspects as it influences the cohesiveness behavior of sediments and
46 in ecology aspects for mapping habitats or evaluating hydration stress conditions
47 for example. Microphytobenthos (MPB), a photosynthetic benthic microalga, is in
48 partly responsible of the bio-stabilization process. Consequently, mapping of MPB
49 biomass is important for fully assessing the reworking sediment processes.

50 Remote Sensing is now considered as an useful tool for mapping physical prop-
51 erties of sediments (Rainey et al. 2003; Deronde et al. 2008; Adam et al. 2010;
52 Small et al. 2009; Verpoorter et al. 2014) or biophysical properties (Combe et al.
53 2005; Méléder et al. 2010; Launeau et al. 2018). While remote sensing tool presents
54 the advantage to get a synoptic overview of coastal zones with wide swaths, scenes
55 analyzed for mapping are only restricted to few millimeters of sediment surfaces
56 limiting our understanding of sedimentary processes over time to approximately one
57 or several tidal cycles. Nevertheless several processes such as erosion, stabilization,
58 bio-stabilization, transport and deposition can be assessed over time using imagery

Table 1 Spatial and spectral resolutions of sensors used

Hypex 1600 VNIR		SPOT 6/7		S2	
Spatial resolution (m/pixel)	Spectral bands	Spatial resolution (m/pixel)	Spectral bands (nm)	Spatial resolution (m/pixel)	Spectral bands (nm)
0.7	160 bands between 400 and 1000 nm	6	450–520	10	425–555
			530–590		525–595
			625–695		635–695
			760–890		
		1.5	450–745		727–957

at high spatial resolution (SPOT 6/7, 1.5 m/pixel; Sentinel-2 (S2), ~10 m/pixel). The high spatial SPOT 6/7 multispectral satellite is well designed for that purpose. Although SPOT 6/7 multispectral data enable us to precisely map sedimentary and geomorphic structures (e.g. ridges and runnels, mega-ripples, spit, waterline, shoreline, etc.) at various scales of observation, S2 sensor presents the benefit to have a very high repetitive time over the same scene (3–5 days with the S2A and S2B configuration) allowing to generate homogenous time-series. Multispectral data contain few spectral channels (Table 1) compared to hyperspectral data. When spatial resolution of hyperspectral sensors which are generally mounted onboard an aircraft are maximized, hyperspectral images are well-suited for deriving accurate quantitative maps with respect to spectral models used. It is explain by the fact that hyperspectral sensor is the only one allowing the collection of a contiguous reflectance spectrum per pixel in numerous narrow spectral bands. In this study, we proposed to use the synergy between various optical sensors operated onboard satellites (SPOT 6/7 and S2) and airborne platforms (HySpex 1600 VNIR) for mapping sedimentary parameters variability at low tide and thus during the same day (09-21-2016).

The main objective of this study consists in 1) mapping sediment properties by tacking the advantage of the synergy between remotely sensed data (SPOT 6/7, S2, and hyperspectral image termed HySpex) and LiDAR data. Bedform dynamics will be examined by calculating topography changes with two sets of LiDAR data between 2013 and 2016.

2 Study Site

The Authie estuary (50° 22' 25" N, 1° 37' 52" E, Fig. 1a) is located on the French coast of the Eastern English Channel (EEC), and it is characterized by a semi-diurnal and macrotidal regime. The Authie estuary is a “picard” type-estuary according to the Briquet et al. (1930) study. The Authie estuary forms a shallow elongated embayment oriented SE/NW (N120). It is affected by a general sandfilling and it presents a typical geomorphology structures, in accretion to the south zone such as ‘la pointe

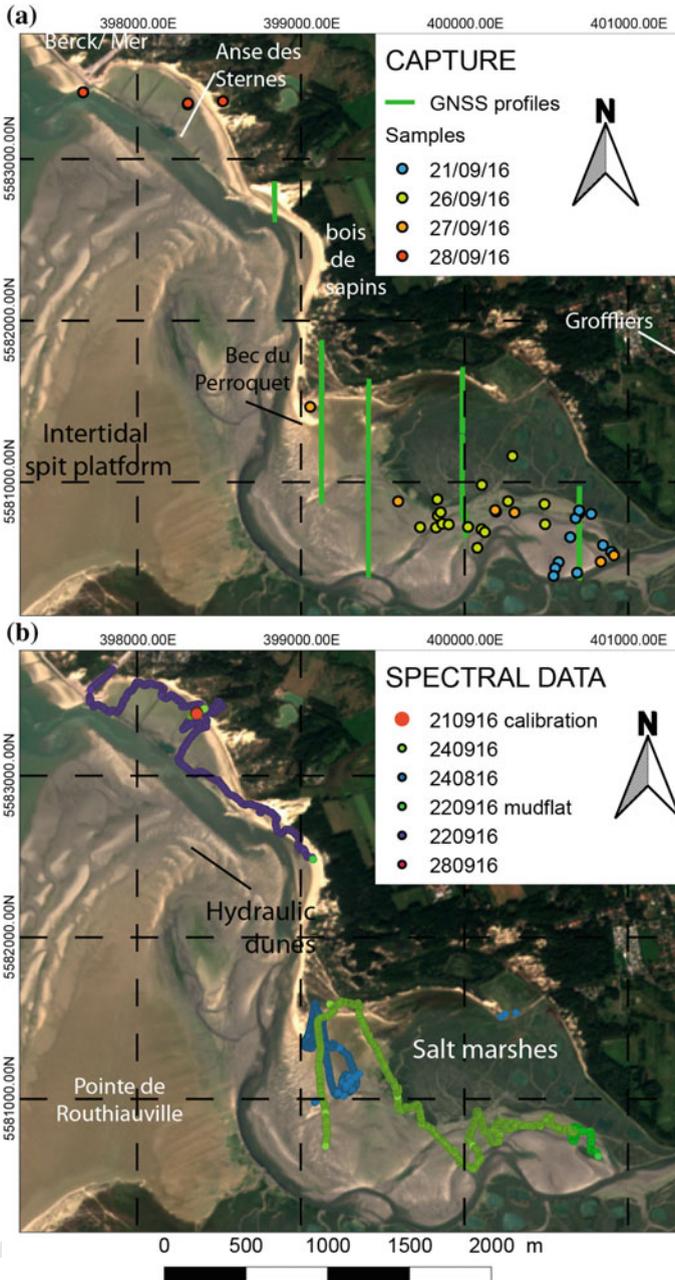


Fig. 1 In situ investigations in the Authie Bay. **a** Sediment samples and GNSS cross-shore profiles, **b** radiometric measurements acquired between the 21th of September 2016 and the 21st of September 2016

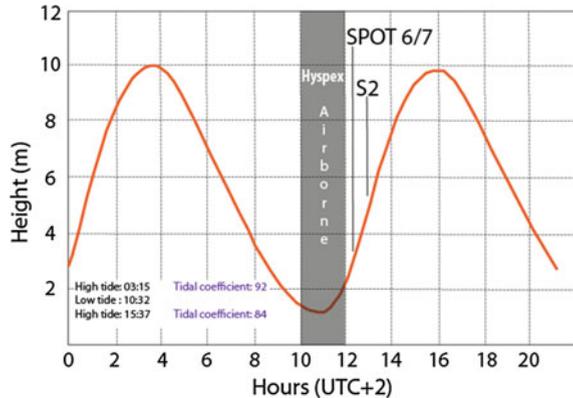
87 de Routhiauville' spit which reflect a northward progradation (~6.1 m/an between
88 2005 and 2012; Dobroniak 2005; Sogreah 2009; Hesp et al. 2016) in contrast with
89 the northern part which is affected by a local erosion. The Authie sea-chenal deviates
90 to the North as a response of the spit accretion and the sandfilling. As a consequence,
91 the northern coastal line is affected by erosion of the secondary sediment cells such
92 as 'la Grande dune' (sand-dune), 'Bois de Sapins' (pine forest) and the 'Anse des
93 Sternes' (a well-protected area) with an erosion rate of 50.000 m³/an (Sogreah 2009).
94 All reworked sand sediments participate to the general infilling of the inner bay. In
95 the medium estuary, a secondary spit structure (a cusplate foreland) was in accretion,
96 termed the 'Beck du Perroquet' testifies to this sandy transport until the salt-marches.
97 It is a southward prograding system (Hesp et al. 2016). The sediment coverage is
98 highly controlled by the tidal conditions with a general decrease in particle-size
99 from the marine estuary to the low estuary. Along the marine side, the morphology is
100 marked by the presence of ridges and runnels (sedimentary bars) which is a particular
101 characteristic of the EEC macrotidal coastal system. It is composed by coarse to
102 medium sizes of sand-deposits. Upper and middle part of the estuary are mainly
103 composed by intertidal sand flats and sandy bars ranging from fine sizes (D50 =
104 0.1 mm) to medium sizes (D50 = 0.6 mm) including a shelly debris component.
105 Low estuary is mainly composed by silty sediments. Mudflats and salt-marches
106 occur mainly in the inner estuary (Deloffre et al. 2007) as well as in the Anse des
107 Sternes as these zones are relatively well-protected from wave-current by breakwater
108 and groynes. Whatever the sites considered MPB, mainly diatoms, can form biofilm
109 patches on muddy sediments. Due to the described overall sandy-transport pattern,
110 it is not rare that during a tidal cycle some sand sediment might recover the existing
111 muddy surfaces within the salt marshes and mudflats.

112 3 Materials and Methods

113 As mentioned before, here it is proposed to map sediment parameters by integrating
114 the synergy between a MultiSpectral Instrument (MSI) on Sentinel-2 (S2) platform,
115 a high spatial resolution image (SPOT 6/7), high spectral resolution images called
116 hyperspectral images (HySpex) and LiDAR topographic data. In the framework of
117 the 'Défi littoral OMPBI' project the Authie Bay was imaged on September 21th
118 2016 between 9:07 and 09:58 (UTC) by a hyperspectral/LiDAR campaign at low
119 tide.

120 Sediments samples (Fig. 1a) and cross-shore elevation profiles were also col-
121 lected using a GNSS (Leica TPS Syst1200) with vertical and horizontal accuracy of
122 ± 2.5 cm and ± 1.5 cm respectively (Fig. 1a, in green line) at the same day for further
123 assessments and validations. Radiometric measurements were also performed using
124 an ASD Fieldspec 4 FR4 [2500 spectral bands]) in order to calibrate hyperspectral
125 images and also have a reliable atmospheric correction (Fig. 1b, red spot). The in situ
126 survey was conducted during one week, where further samples and radiometric data
127 were collected (Fig. 1). A total of 65 surficial sediment samples were collected across

Fig. 2 Tidal condition and airborne-satellites overpass configuration in 21st of September 2016



128 the intertidal and supratidal domains during one week and at low tide (Fig. 1a).
 129 Conditions for further investigation in grain-size, chlorophyll-a concentration (10
 130 samples).

131 In situ and laboratory dataset were used for validation of spectral models. Then
 132 sediment samples were analyzed under laboratory conditions. Chlorophyll-a contents
 133 were measured using a spectrophotometer on a set of 10 samples acquired during the
 134 overflight. Grain-sizes analyses were performed using LS230 laser particle-sizer (©
 135 Beckman Coulter). Grain size parameters (mean, sorting and skewness), sedimentary
 136 facies and classes were calculated using the GRADISTAT software (Blott and Pye
 137 2001) following the Folk and Ward (1957) methods.

138 Multispectral remotely sensed data (SPOT 6/7 and S2) were acquired more or less
 139 30–60 min after the airborne overpass (Fig. 2). While LiDAR data were acquired
 140 simultaneously with the hyperspectral images (HySpex'sensor) we can assume that
 141 topography and main geomorphic structures did not affected by significant changes
 142 from an image to another one, at least in the inner estuary. Such as multi-sensors
 143 configuration is judged to be really rare and it offers us the opportunity to map
 144 sedimentary facies under topographic controlled during the immersion time.

145 SPOT6/7 image was ordered in the frame of the GEOSUD program. Initially
 146 the multispectral bands (6 m/pixel) were pan-sharpened to the panchromatic one
 147 (1.5 m/pixel) for enhancing the spatial resolution of all spectral bands to 1.5 m/pixel
 148 using the Gram –Schmidt pan-sharpening technique (Laben and Bower 2000).
 149 Then, raw images were calibrated and corrected from atmospheric effects using
 150 the FLAASH method. With a spatial resolution of 1.5 m/pixel (Table 1), the spatial
 151 resolution is sufficient to discriminate accurately sedimentary structures of bedforms
 152 such as sedimentary bars, mega-ripples, creeks, sea-channel, etc.

153 Sentinel-2 sensor was operated belong ESA (European Spatial Agency) and deliv-
 154 ered throughout the THEIA platform into a L2A format following the MAJA method
 155 (Hagolle et al. 2010). MAJA is an evolution of MACCS (Multi-sensor Atmospheric
 156 Correction and Cloud Screening), in which a couple of methods inspired by ATCOR

157 software have been added. With a spatial resolution of 10–20 m per pixel, the resolution
 158 is not sufficient to discriminate accurately all sedimentary structures at fine-scale
 159 such as some of the mega-ripples. However, S2 is well-suited for mapping meso-
 160 scale structures such as sandy spit, main channel, ridges and runnels, some of the
 161 mega-ripples, etc.

162 Hyperspectral images were acquired with a HySpex Visible Near InfraRed
 163 (VNIR) 1600 camera having a spectral resolution of 4.5 nm in 160 spectral channels
 164 between 400 and 1000 nm. Simultaneously, topographic data were acquired with the
 165 LiDAR platform of Nantes-Rennes universities with a spatial resolution of 1 m/pixel.
 166 Fly conditions was set-up for maximizing the spatial resolution of hyperspectral data
 167 at 0.7 m per pixel. Atmosphere correction was performed with the ATCOR 4/ MOD-
 168 TRAN. A Minimum Noise Fraction (MNF) transformation (Green et al. 1988) was
 169 applied to images to reduce noise from signal.

170 A digital terrain model (DTM) was generated with a spatial resolution of 1 m per
 171 pixel (Fig. 3). In addition, GNSS data (2273 ground control points) were used for
 172 minimizing error in areas covered dense vegetation. DTM correction was performed
 173 based on a linear regression approach between LiDAR data and GNSS ground control

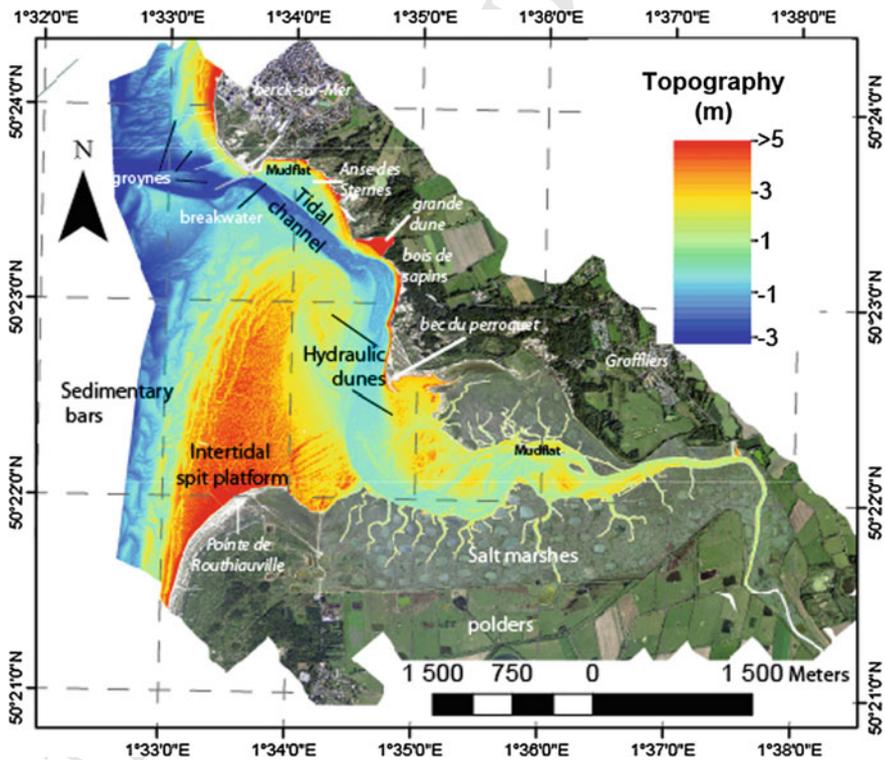


Fig. 3 Morphology map of the Authie estuary generated from LiDAR data, 21st of September 2016

174 points ($y = 1.005 - 0.39x$, $R^2 = 0.99$, $RMSE = 0.386$). Figure 3 displays the bay
175 morphology including ridge and runnels, crest of sedimentary bars, hydraulic dunes,
176 salt marshes, etc. In addition, in the framework of the SNO DYNALIT previous a
177 LiDAR topographic data (ALS60 LiDAR, © Leica Geosystems) was acquired in
178 December 2013 by the Circle operational teams. The planimetric position accuracy
179 of the data points range from ± 0.10 to 0.17 m with a vertical accuracy ± 0.10 m
180 as verified by several ground control points. In this study we only compare the
181 2016 LiDAR data to the 2013 one. Topographic discrepancies were computed for
182 interpreting the morpho-sedimentary dynamics and characteristics of the sea-bed
183 estuary.

184 With the intention to derive maps of sediment parameters, several spectral mod-
185 els were used according to the spectral resolution. When mapped with multispectral
186 sensors, biomass is usually expressed in NDVI (Normalizes Differential Vegetation
187 Index, Tucker 1979) values. Another spectral index called MPBI index was proposed
188 by M el eder et al. (2010) to distinguish MPB from macrophytes. In this study, two
189 methods were tested. Combe et al. (2005) developed a non-linear unmixing model for
190 hyperspectral data based on MGM (Modified Gaussian Model) continuum removal.
191 Even if this model is recognized to be well-appropriate for very high spectral reso-
192 lution data, we tested a similar approach designed for multispectral data. Basically
193 the developed approach stills the same except that we didn't perform any continuum
194 removal prior to derive biomass and it is based on a linear unmixing approach. For
195 simplicity, this method was called LMM for Linear Mixing Method. Compared to
196 spectral indices methods, LMM has the advantage to integrate sandy and muddy
197 surfaces into the mixing itself.

198 However, when mapped with hyperspectral resolution, advanced spectral model
199 may be used allowing to analyze the full shape of continuous reflectance spectra
200 and narrow absorption bands characteristics with respect to different MPB groups
201 (Combe et al. 2005). In the framework of the OMPBI project, a radiative transfer
202 algorithm called MPBOM (MicroPhytobenthos Optical Model) was applied with
203 hyperspectral data (Launeau et al. 2018). While many biomass products could derived
204 from this model, here the latter was only used to retrieve the biofilm absorption
205 coefficient α which was linearly related to biomasses. Then, MPBOM outputs were
206 briefly compared to the unmixing algorithm outputs.

207 With the intention to compare maps of grain-size and water content param-
208 eters for each of sensors used, we decided to apply a simple multispectral algo-
209 rithm (see *more detailed in appendices*) based on spectral features and in situ mea-
210 surements of collocated sediment surfaces. Basically, algorithms consisted in con-
211 fronting in situ measurement (mean grain-size) to reflectance value of the images. For
212 such as comparison, hyperspectral images were resampled to the SPOT 6/7 spectral
213 resolution.

4 Results

4.1 Morphology of the Seabed

Surface elevation evolution can be retrieved using LiDAR data acquired at different periods. Topographic differences enable us to quantify precisely the relative sedimentary budget trend. Crapoulet (2015) calculated a positive budget for the overall Authie estuary from LiDAR data acquired between 2008 and 2011 about $827,162 \text{ m}^3$ ($+0.08 \text{ m}^3/\text{m}^2$). Conversely, they calculated a negative budget for the overall Authie estuary from LiDAR data acquired between 2011 and 2013 about $-569,529 \text{ m}^3$ ($-0.05 \text{ m}^3/\text{m}^2$). As a consequence, between 2008 and 2013 there is a volumetric gain of $+111,191 \text{ m}^3$ ($+0.01 \text{ m}^3/\text{m}^2$).

We propose here to revise the budget number by integrating new LiDAR data acquired during the OMPBI survey from 2013 to 2016. Basically DTMs relative map differences were computed exhibiting a sea-bed dynamics allowing to identify erosion (blue-yellow-orange) and accretion (in red) zones. The relative sedimentary budget remains negative over the overall estuary (Mean = $-0.41 \pm 0.859 \text{ m}^3/\text{m}^2$) which is consistent with Crapoulet (2015) study. Indeed, a negative budget for the overall Authie estuary from LiDAR data acquired between 2011 and 2013 about $-219,239 \text{ m}^3$ ($-0.04 \text{ m}^3/\text{m}^2$). As a consequence, between 2008 and 2016 there is a volumetric loss of $-108,048 \text{ m}^3$ ($-0.02 \text{ m}^3/\text{m}^2$). Consequently the net changes in elevation surface between 2008 and 2016 reveals a significant loss of volume on the overall estuary.

Because ridge and runnels (sedimentary bars) and mega-ripples systems are highly dynamics from a tidal cycle to another one, they are not representative of the sedimentary budget for such a long period ~ 3 years. Intentionally we do not consider them into our sedimentary budget and we restricted our calculations to several representative zones within the inner estuary ('Bois de Sapins', mudflats, 'Anse des Sternes', 'Bec du Perroquet', middle estuary and spit platform).

Observations (Fig. 4) are consistent with literature as the 'Bois de Sapins' is affected by strong erosion (blue color) with a decrease of elevation ranging from -9 to 3 m (Mean = $-1.2 \pm 1.3 \text{ m}^3/\text{m}^2$). This is explained by the fact that the estuary mouth is largely blocked by the prominent sand spit platform, and this has forced the Authie channel of the estuary to flow along the northern shore. As the consequence, this zone concentrates a maximum of energy. Despite that the anthropic area called 'Anse des Sternes' is well protected from waves and currents with breakwater and groynes structures, the sedimentary budget is slightly negative with an elevation ranging from -6 to 0.8 m (Mean = $-0.5 \pm 0.2 \text{ m}^3/\text{m}^2$). The eastern part of the 'Bec du Perroquet' presents slight negative elevation variations with an elevation surface ranging from -2.3 to 1.5 m (Mean = $-0.23 \pm 0.27 \text{ m}^3/\text{m}^2$). Natural mudflats located along the northern salt-marches are affected by a slight erosion (yellow color) with a decrease of elevation ranging from -2 to 0.65 m (Mean = $-0.3 \pm 0.3 \text{ m}^3/\text{m}^2$). The spit platform budget is slightly negative with an elevation ranging from -1.6 to 1.0 m (Mean = $-0.41 \pm 0.01 \text{ m}^3/\text{m}^2$). Negative budgets observed in mudflats, spit and

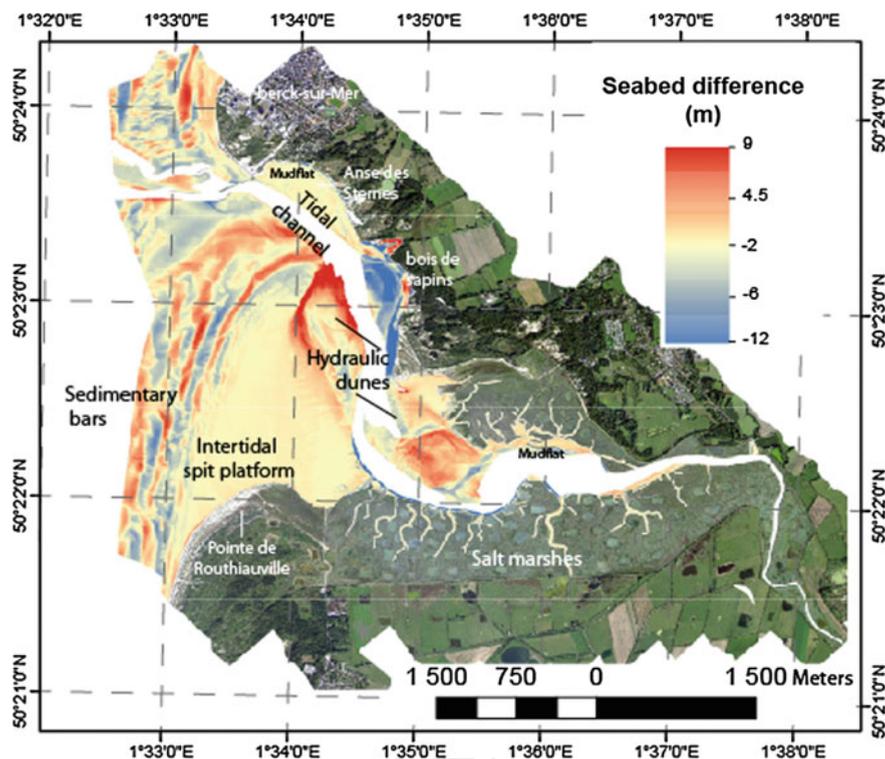


Fig. 4 Net changes in surface elevation between December 2013 and September 2016 based on LiDAR topographic data

256 anywhere else can be explained by the fact that the sediment supply may decrease
 257 over time explaining that there is less deposits or that the sediment compaction
 258 processes might more efficient.

259 While the overall estuary presents a negative budget, several zones are clearly
 260 affected by accretion such as the middle estuary and the north of the backside part
 261 of the spit. In the middle estuary, there is a positive elevation (in red) variation
 262 ranging from -1.3 to 1.3 m (Mean = 0.25 ± 0.30 m³/m²). This positive budget
 263 can be explained by the fact that the allochthonous sediments coming from the
 264 maritime side and autochthonous sediments coming from erosive sites may increase
 265 over time due to high sedimentary transit during the flood-tide and thus until their
 266 deposits. To conclude some sediment accumulations occurred downdrift to the south,
 267 but only took place at middle estuary. The backside of the spit platform budget
 268 presents a positive budget (in red) mainly in the northern part with an elevation
 269 ranging from -0.83 to 2.8 m (Mean = 1.0 ± 0.6 m³/m²). This may be explained
 270 by the fact that the spit is prograding toward the N–W part of the bay. In addition,
 271 reworking sediments coming from the erosion sites may contribute of some sediment
 272 accumulations downdrift to the south, but only took place at tip of northern spit.

273 Finally, regarding the supratidal domain, ‘la Grande Dune’, a parabolic dune seems
274 to be in accretion at the top due to Aeolian deflation inputs which may coming
275 from intertidal sandy bars, generally drier than the rest of intertidal sediments. The
276 bottom part of this dune is in erosion. With a ranging from -10 to 9 m (Mean = $-$
277 0.65 ± 2.40 m³/m²) the overall budget is slightly negative.

278 4.2 MPB Spatial Distribution Mapping

279 As mentioned before a LMM spectral model was applied at multispectral resolution
280 sensors allowing to compare biomass output products together (Fig. 5). Whatever the
281 sensor used MPB generally occurs over fine sediment substrates such as very fine-
282 sands or silts. High MPB concentrations are mainly present on mudflat sediments
283 (‘Anse des Sternes’ and along salt-marshes). In details, MPB was locally observed
284 at the bottom of mega-ripples with respect to the spatial resolution of the sensors.
285 This is explained by the fact that grain-size may slightly decrease and that moisture
286 content is generally higher compare to mega-ripple crests. Surprisingly, MPB also
287 occurs on most part of the spit. Despite that the spit structure is mainly composed
288 by sandy sediments (135 ± 40 μ m; HySpex), it is possible to observed a very thin
289 layer composed by muddy sediment covering the sand substrates. This might appear
290 when tide energy becomes lower during low tide. However, it is well-known that MPB
291 distributions are not only a function of the sediment cover-type (sand or mud), it may
292 also vary in time and space as a function of numerous factors like: time of emersion,
293 solar energy, and nutrients inputs, etc. Because images were acquired at various times
294 of acquisitions at the same day, sensors synergy offers us a unique opportunity for
295 mapping the evolution of the MPB distribution during a tidal cycle especially at the
296 emersion time. Figure 5 exhibits strong spatial variation in biomass concentrations
297 which means that MPB migration is highly dynamic at short time during aerial
298 exposure. It seems that the MPB concentration spatially decrease from 10:00 to
299 13:00 UTC+2 which means that the maximum of MPB production is observed with
300 a mean value of 27 ± 19 mg.m⁻² during the HySpex overpass (Fig. 5a) while the
301 minimum of MPB production (Fig. 5b) is observe during the S2 overpass (Fig. 5c)
302 with a mean value of 17 ± 13 mg m⁻². Intermediate values are observed during the
303 SPOT 6/7 overpass (Fig. 5b) with a mean value of 21 ± 18 mg m⁻². Note that the
304 MPB variability was correlated to the PAR (Photosynthetically Active Radiation)
305 values. Although these general patterns reveal a decrease in biomass coverages, in
306 details some patches do not follow this trend. Clearly the S2 image reveals some
307 patches with higher chlorophyll-a content than the SPOT6/7 or HySpex ones.

308 Although MPB production seems to be correlate to the solar radiation changes as a
309 first order, other parameters may have a significant influence on it. Indeed, it is known
310 that MPB migration may be influenced by the tidal range. For example, we denote
311 that during the S2 overpass, the flood-tide coming back in the western part of the
312 upper estuary, as a consequence it may induce an impact in our estimates. On the other
313 hand, spatial resolution may also have a significant impact on MPB estimates due to

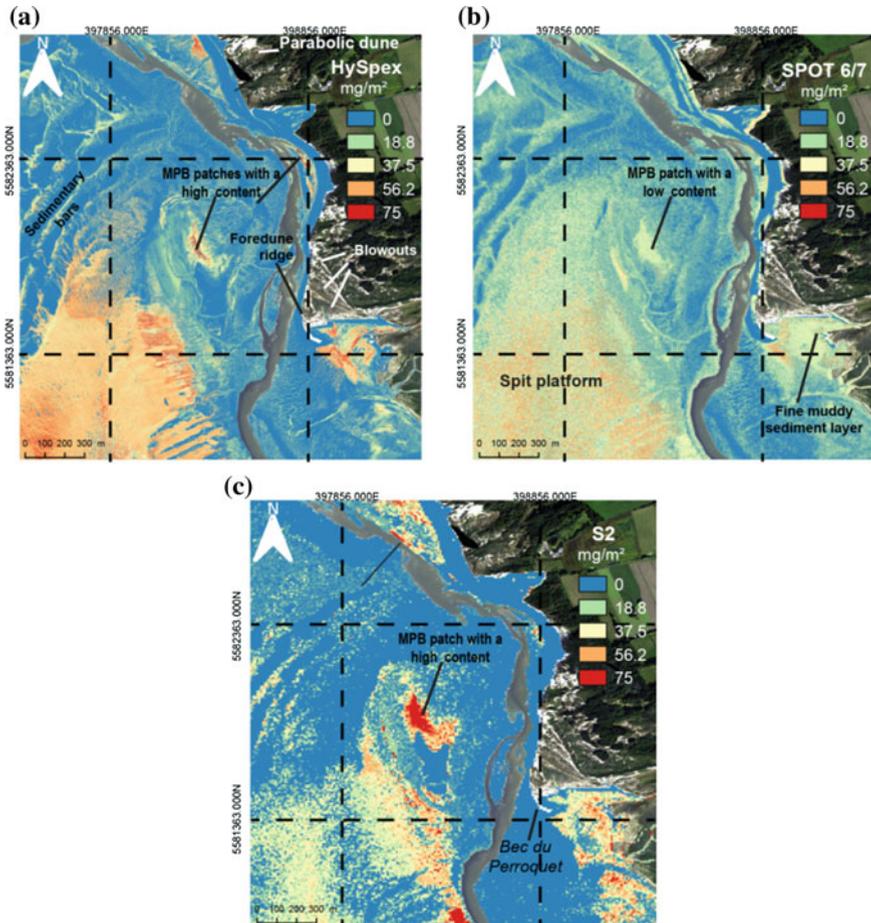


Fig. 5 Maps representing the net primary production (in mg m^{-2}) at low tide (09/21/2016) from **a** HySpex VNIR; **b** SPOT 6/7 and **c** S2 data and by applying the Liner Mixing Model (LMM)

314 non-linear mixing between natural surfaces within a pixel. Because, S2' resolution
 315 is coarser than SPOT 6/7 (with a factor ~ 3) and HySpex ones (with a factor ~ 14),
 316 the probability that MPB covers are mixed with abiotic surfaces (e.g. sand or mud)
 317 is pretty high. As a result, biomasses may be underestimated. It is one of the reasons
 318 why more advanced methods such as non-linear unmixing model (Combe et al. 2005)
 319 or none-scale dependent approaches such as the MPBOM (Kazemipour et al. 2012)
 320 still desirables. However, those methods cannot be applied on multispectral data.

321 MPBOM was applied to hyperspectral HySpex images in the framework of the
 322 'Défi littoral OMPBI' project. Among biomass products (coefficient of absorption,
 323 peak of the Chl-a, biomass of the MPB) the distribution of the absorption coefficient
 324 peak at 673 nm was analyzed. Figure 6 display the total chlorophyll-a distribution

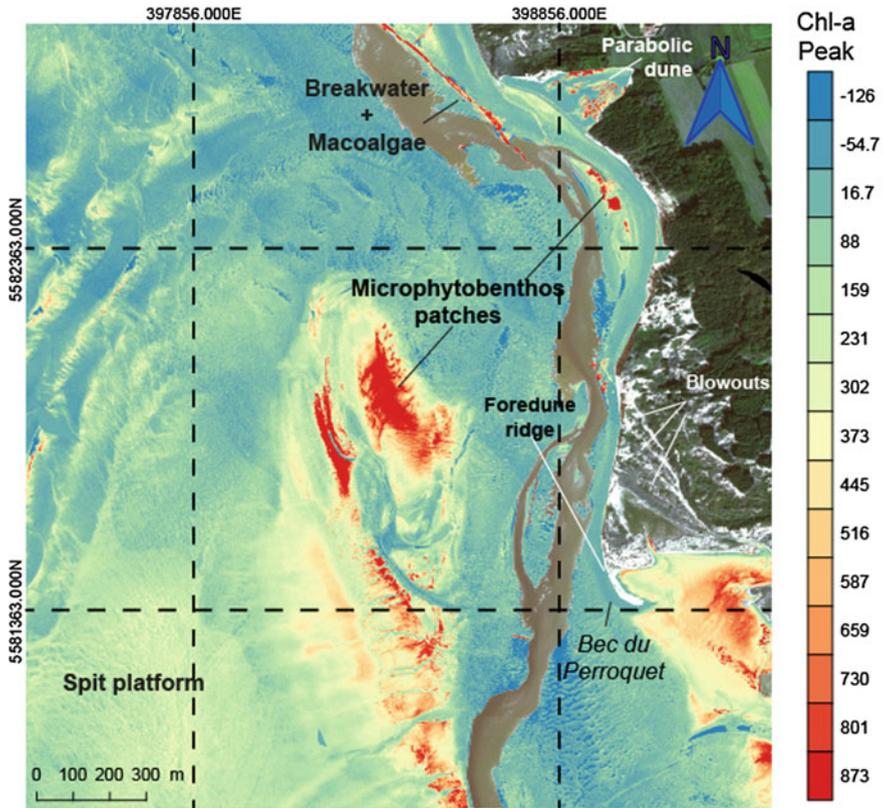


Fig. 6 Map representing the peak of the absorption coefficient at 673 nm during low tide (09/21/2016) from HySpex VNIR data and by applying the MPBOM in the framework of the 'Défi littoral OMPBI' project

325 deriving from the absorption coefficient peak at 673 nm. High absorption contents are
 326 observed on the breakwater. High chlorophyll contents composed by microphyto-
 327 benthos are generally observed in the backside of the spit platform and muddy sites.
 328 As mentioned before, during the airborne campaign 10 samples were collected for
 329 calculating the chlorophyll-a concentration and compared them at the pixel location
 330 of the HySpex data. MPBOM pixels are positively correlated with these samples (R^2
 331 = 0.57). However, for better assessment more samples are needed. MPBOM product
 332 (Fig. 6) exhibits clearly strong discrepancies with the LMM (Fig. 5a), especially over
 333 the spit whereas similar trend is observed with S2 data especially over some patches
 334 (Fig. 5c).

335 Because MPBOM take into account non-linear mixing effects, we can assume that
 336 non-linear effects may induce an overestimation of the biomass with multispectral
 337 models used here.

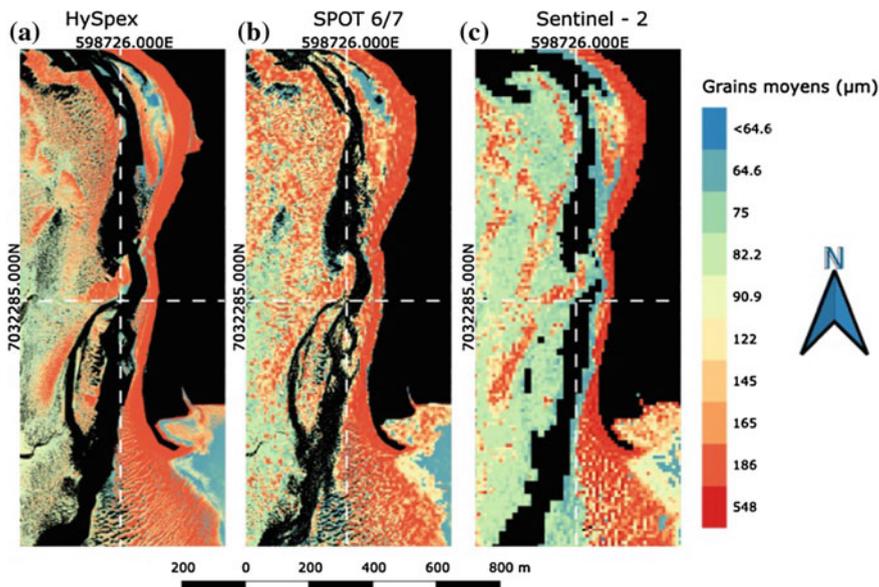


Fig. 7 Maps representing the grain-size distribution (in μm) at low tide (09/21/2016) from A) HySpex VNIR; B) SPOT 6/7 and C) S2 data and by applying experimental spectral model

4.3 Physical Properties of Surficial Sediments

As mentioned before an experimental spectral model was applied at multispectral resolution sensors allowing the comparison of physical properties (grain-size and moisture content) products together. Whatever the sensor used, grain-size (Fig. 7) and moisture content (Fig. 9) display similar patterns with a high spatial correlation.

Grain-size product validations were performed using true field data as reference. During the airborne campaign among the overall collected samples, 26 samples were analyzed using Coulter LS230 laser particle-sizer for calculating the mean grain-size and compared them at the pixel location of the satellite images. RMSE calculations ($\text{RMSE}_{\text{HySpex}} = 52 \mu\text{m}$; $\text{RMSE}_{\text{SPOT 6/7}} = 40 \mu\text{m}$ and $\text{RMSE}_{\text{S2}} = 44 \mu\text{m}$) reveals that the range of variation between pixel values and true field data is quite low providing reasonable estimates for differencing sediment surface fraction sizes.

In Fig. 7, coarse sediments (red color) are observed on beaches, on the crest of mega-ripples, on sediment sandy bars, spit, and dunes where topography values are relative higher than the rest of the slikke (green-blue color). Locally, coarse sediments are present in shell beds deposit. The intertidal domain is mostly composed by fine sand and very fine sand sediments ($\sim 85\%$). Silty sediments are particularly observed on mudflat (blue color) where hydrodynamic conditions are low in energy. Mean grain-size of the Authie Bay reaches an average of $120 \pm 50 \mu\text{m}$ of $170 \pm 70 \mu\text{m}$ of $160 \pm 70 \mu\text{m}$ at the HySpex, SPOT 6/7 and S2 resolutions. S2 image displays a large amount of fine-sediments pixels (in green) in comparison with the HySpex

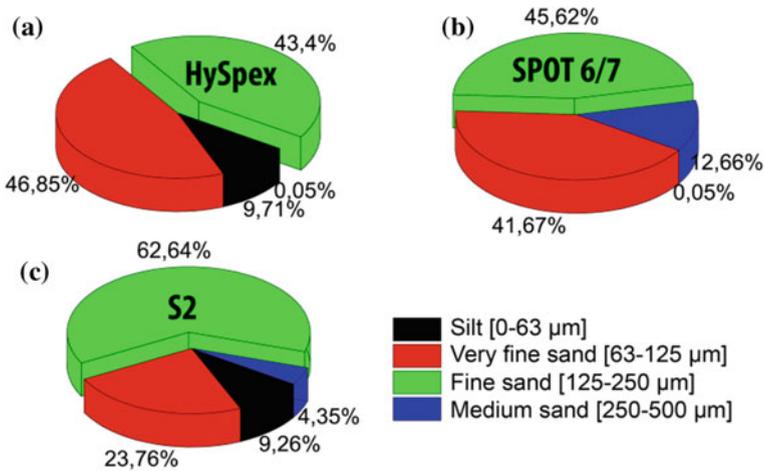


Fig. 8 Pie diagrams representing the grain-size distribution (in %) for each sensor: A) HySpex VNIR; B) SPOT 6/7 and C) S2 images

359 image which display coarser sediment pixels (in red). The SPOT 6/7 image displays
 360 intermediate values (in orange). With respect of the spatial resolution used, grain-size
 361 products display some major discrepancies. Indeed the coarser the spatial resolution is,
 362 the coarser the grains-size is (Fine sand + medium sand).

363 Whatever the sensor used (Fig. 8), fine sand sediments represent the grain-size
 364 fraction which is the more abundant (HySpex: 43%, SPOT 6/7: 45% and S2: 62%)
 365 followed by very fine sands (HySpex: 46%, SPOT 6/7: 41% and S2: 23%) silts
 366 (HySpex: 9%, SPOT 6/7: 45% and S2: 62%) and medium sands (SPOT 6/7: 12% and
 367 S2: 9%).

368 Accordingly, grain-size products are impacted by the spatial resolution of the sensors
 369 which might lead to an overestimate or an underestimate in the final products. Again,
 370 this is explained by the non-linear unmixing effect which is not taken into account with
 371 experimental spectral models developed. Contrary to microalgae surfaces which are
 372 highly dynamic over time, grain-size fractions should not exhibit strong changes
 373 during the low tide, even slightly reworking may locally appear (e.g. grazing, feeding).
 374 As a consequence, we can assert that spectral resolutions and scale factors are
 375 respectively main factors explaining those discrepancies.

376 In Fig. 9, dry sediments (red color) are observed on beaches, on crests of mega-
 377 ripples, on sediment sandy bars, spits, and dunes where topography is relatively higher
 378 than the rest of the slikke (green-blue color). Very moist sediments are observed in
 379 mudflats. It is not rare to observe a thin surficial water layer over sediment substrates.
 380 Water content is widely influenced by the porosity of sedimentary structures as well
 381 as time of aerial exposure during a tidal cycle. It is clear that grain-size patterns are
 382 spatially correlated to water content patterns. This is simply explained by the fact that
 383 coarser sediments are drier than silty sediments due to the higher efficiency in the
 384 percolation process. Water content products are also influenced by spectral resolutions of

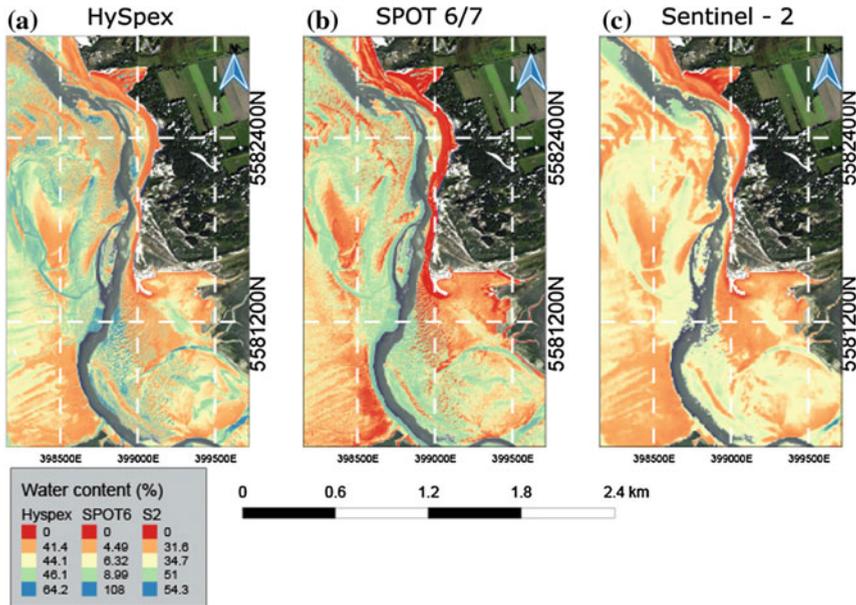


Fig. 9 Maps representing water content (in %) at low tide (09/21/2016) from **a** HySpex VNIR; **b** SPOT 6/7 and **c** S2 data and by applying experimental spectral model

385 sensors used. Some of the S2 water-pixels seems to be underestimated in comparison
 386 with the HySpex and SPOT 6/7 ones.

387 5 Conclusions and Perspectives

388 The following conclusions may be made from this study:

- 389 i. Net changes in surface elevation between December 2013 and September 2016
 390 based on LiDAR topographic data were computed allowing to revise formal
 391 sedimentary budgets. We demonstrated that the Authie Bay is affected by a
 392 relative loss of volume between 2008 and 2016 with a sediment loss of $-$
 393 $108,048 \text{ m}^3$ ($-0.02 \text{ m}^3/\text{m}^2$) over the overall Authie estuary. Based on LiDAR
 394 data acquired between 2013 and 2016, a significant loss of volume was calculated
 395 about $-219,239 \text{ m}^3$ ($-0.04 \text{ m}^3/\text{m}^2$). To summarize, the sedimentary budget
 396 was highly positive between 2008 and 2011, negative between 2011 and 2013
 397 (Crapoulet 2015) and remains negative between 2013 and 2016 (this study). The
 398 remaining question is what mechanisms control the sediment supply within the
 399 Authie Bay? It might be driven by the fact that there is less sediments inputs
 400 likely due to a decrease of the Holocene stocks in the subtidal domain or it might
 401 be a response of the spit extension which protects the inner estuary from natural

- 402 sediment supply. Consequently the bay infilling might be caused by a decrease
403 of tidal energies or currents within the estuary resulting with an increase of mud
404 deposits along salt marshes, an infilling of the secondary channels across the
405 salt-marches and a salt-marches prograding.
- 406 ii. In details all domains present negative sediment budgets except in the middle
407 estuary and in the tip of the spit. However, it is well-known that the Authie bay
408 is affected by a general infilling while locally some sites appear to be eroded
409 between 2013 and 2016 such as the '*Bois de sapins*' site. However, it is reason-
410 able to mention that such as analyze is relative as the morphology of the bay
411 may drastically change with storm impacts for example. It is difficult to calcu-
412 late an absolute budget at high frequency using few LiDAR data following a
413 longer time serie. In that sense, UAV acquisitions combined to photogrammetry
414 analysis may help to fill the gap of information.
- 415 iii. We demonstrated that is possible to map sediment bio-physical properties using
416 both hyperspectral and multispectral data (HySpex, SPOT 6/7 and S2). We pro-
417 posed to use LMM with multispectral data and MPBOM with hyperspectral data.
418 Whatever the approach used, spatial patterns of bio-physical properties exhib-
419 ited are spatially consistent with field observations and laboratory measurements
420 and our knowledge of sedimentary processes. Whatever the sensor used, grain-
421 size distributions and water content reveal similar patterns. Grain-size and water
422 content are spatially highly correlated with topography. However, further invest-
423 gation still needed for fully understand the role of the geomorphic structures
424 on sediment deposits as well as the net primary production. As perspectives,
425 grain-size maps may be used as inputs of transport models (like GSTA, 3D
426 models, etc.) and thus for computing transport trends. Moisture content is also
427 an important parameter to consider as the latter may be used for improving our
428 knowledge about the cohesive behavior, the role of the Aeolian deflation over
429 the intertidal sandy bars towards dunes in accretion for instance, etc. Finally, it
430 is crucial to assess the distribution of microphytobenthos as the latter influence
431 the bio-stabilization processes and thus the reworking of intertidal sediments.
- 432 iv. Regarding the biomass products, data generated from LMM may be overesti-
433 mated particularly on the spit compared to MPBOM product. As a consequence,
434 spectral algorithm choice differs from the sensor used. This may be explained
435 by the fact that both non-linear mixing effects and scale factors are not taking
436 into account with multispectral models.
- 437 v. Multispectral experimental models were developed with the intention to com-
438 pare images each other. Developed algorithms were based on a learning dataset
439 approach by confronting spectral behaviors directly to physical properties of
440 the sampled sediments. Although we do not demonstrated that these algorithms
441 are not reproducible to other coastal domains, in the frame of this study they
442 remain useful for improving our understanding of the estuary evolution during a
443 low tide cycle. In that sense, the synergy between various sensors presents some
444 interesting advantages. For instance, it seems possible to predict net primary
445 budget over time by analyzing data at various time of acquisitions. However,

446 with the intention to determine the impact of the non-linear mixing effect on
447 products more similar assessments are needed.

448 vi. To conclude, when a long time serie is available such as the multispectral S2-MSI
449 or Landsat-OLI images, it becomes possible to better understand some processes
450 like erosion and sandfilling interactions over time. With a high revisiting time
451 over the same scene (5–10 days), Sentinel-2 data would be particularly well-
452 adapted for assessing effects of seasonal variation or storm events and their
453 respective geomorphic variations. L8 have a temporal resolution of 16 days per
454 scene acquired. Note that Landsat serie is known to be the longest time-serie
455 available in remote sensing starting from 1972 with Landsat-1 platform to now
456 with Landsat-8 platform. It would be interesting to process such as data for
457 better assessing the infilling of picard estuaries.

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469 Nord-Pas de Calais and by the French Centre National de la Recherche Scientifique (CNRS).

470 **Appendice 1: Grain-Size Algorithm Used for Mapping**

471 With the intention to derive map grain-size properties from remotely sensed data,
472 a spectral algorithm was computed based on the matching technique between field
473 sampling (grain-size value) and spectral information (reflectance values) for each of
474 sampling pixels. Mean grain-size is computed such as (Eqs. 1 and 2):

$$476 \quad Coarse = 215.30 * \log\left(\frac{R_{red} - R_{green}}{R_{NIR} - R_{bleue}}\right) + 290.69 \quad (1)$$

$$478 \quad Fine = \left| -121.17 * \left(\frac{R_{red} - R_{green}}{R_{NIR} - R_{bleue}}\right) + 290.69 \right| \quad (2)$$

479 where:

480 R_X Reflectance values for a spectral channel.

481 These two resulting maps (fine fraction and coarse fraction maps) were merges
482 for obtaining a mean grain-size map. Merging criteria was defined by the following

483 ratio (Eq. 3):

$$486 \quad \text{Ratio} = \frac{R_{red} - R_{green}}{R_{NIR} - R_{red}} > X \quad (3)$$

486 where:

487 X Muddy/Sandy ratio, empirically fixed by the operator. (for this study, X range
488 from 0.20 to 0.25).

489 **Appendice 2: Water Content Algorithm Used for Mapping**

490 With the intention to derive maps of water content from remotely sensed data, a
491 spectral algorithm was computed based on laboratory measurements following the
492 Verpoorter et al. (2014) approach. It consists in measuring reflectance response dur-
493 ing dehydration process from representative sediments. Then, the resulting spectral
494 behavior was modelled using an exponential regression function by selecting the red
495 wavelength channel which is known for responding well to the water content changes
496 ($R^2 = 0.77$; $P < 0.0001$). Finally, the resulting model was applied at the remotely
497 sensed images following the Eq. (4):

$$498 \quad [\text{H}_2\text{O}] \% = 87.42 * \exp\left(-R_{red}/0.07\right) - 3.61 \quad (4)$$

500 **References**

- 501 Adam, S., De Backer, A., De Wever, A., Sabbe, K., Toorman, E. A., Vincx, M., et al. (2010). Bio-
502 physical characterization of sediment stability in mudflats using remote sensing: A laboratory
503 experiment. *Continental Shelf Research*. <https://doi.org/10.1016/j.csr.2009.12.008>.
- 504 Briquet, A. Le littoral du nord de la France.
- 505 Blott, S. J., & Pye, K. (2001). GRADISTAT: a grain-size distribution and statistics package for the
506 analysis of unconsolidated sediments. *Earth Surf. Processes Landforms*, 26, 1237–1248.
- 507 Crapoulet, A. (2015). Evolution du trait de côte, bilan sédimentaire et évaluation des zones à
508 risques sur le littoral du Nord-Pas-de-Calais: Analyse multiéchelles par LiDAR aéroporté. *These*
509 *de doctorat en Géographie Physique, Wimereux: Université du Littoral Côte d'Opale*.
- 510 Combe, J. P., Launeau, P., Carrere, V., Despan, D., Meleder, V., Barillé, L., et al. (2005). Mapping
511 microphytobenthos biomass by non-linear inversion of visible-infrared hyperspectral images.
512 *Remote Sensing of Environment*, 98, 371–387. <https://doi.org/10.1016/j.rse.2005.07.010>.
- 513 Deloffre, J., Verney, R., Lafite, R., Lesueur, P., Lesourd, S., & Cundy, A. B. (2007). Sedimentation
514 on intertidal mudflats in the lower part of macrotidal estuaries: Sedimentation rhythms and their
515 preservation. *Marine Geology*, 241, 19–32. <https://doi.org/10.1016/j.margeo.2007.02.011>.
- 516 Deronde, B., Houthuys, R., Henriët, J.-P., & Lancker, V. V. (2008). Monitoring of the sediment
517 dynamics along a sandy shoreline by means of airborne hyperspectral remote sensing and LIDAR:

- 518 A case study in Belgium. *Earth Surface Processes Landforms*, 33, 280–294. <https://doi.org/10.1002/esp.1545>.
- 519
- 520 Dobroniak, C. (2005). Morphological evolution and management proposals in the Authie Estuary, northern France. *Proceeding Dunes and Estuaries*, 2205, 537–545.
- 521
- 522 Folk, R. L., & Ward, W. C. (1957). Brazos river bar: A study in the significance of grain-size parameters. *Journal of Sedimentary Petrology*, 27, 3–26.
- 523
- 524 Green, A. A., Berman, M., Switzer, P., & Craig, M. D. (1988). A transformation for ordering multi-spectral data in terms of images quality with implications for noise removal. *IEEE Transactions on Geoscience and Remote Sensing*, 26(1), 65–74.
- 525
- 526
- 527 Hagolle, O., Huc, M., Villa Pascual, D., & Dedieu, G. (2010). A multi-temporal method for cloud detection, applied to FORMOSAT-2, VEN μ S, LANDSAT and SENTINEL-2 images. *Remote Sensing of Environment*, 114(8), 1747–1755.
- 528
- 529
- 530 Hesp, P. A., Ruz, M.-H., Hequette, A., Marin, D., & Miot da Silva, G. (2016). Geomorphology and dynamics of a travelling cusate foreland, Authie estuary France. *Geomorphology*, 254, 104–120. <https://doi.org/10.1016/j.geomorph.2015.11.015>.
- 531
- 532
- 533 Kazemipour, F., Launeau, P., & Méléder, V. (2012). Microphytobenthos biomass mapping using the optical model of diatom biofilms: Application to hyperspectral images of Bourgneuf Bay. *Remote Sensing of Environment*, 127, 1–13.
- 534
- 535
- 536 Launeau, P., Méléder V., Verpoorter, C., Barillé, L., Kazemipour F., Giraud, M., et al. (2018). Microphytobenthos biomass and diversity mapping at different spatial scales with hyperspectral optical model. *Remote Sensing*, 10(5), 716.
- 537
- 538
- 539 Laben, C. A., & Bower, B. V. (2000). Process for enhancing the spatial resolution of multispectral imagery using pan-sharpening. US Patent 6.011875.
- 540
- 541 Latteux, B. (2001). *Synthèse sur les problématiques et enjeux d'évolution du littoral haut-normand et picard-1*. DDE de la somme: Milieu Physique.
- 542
- 543
- 544 Rainey, M. P., Tyler, A. N., Gilvear, D. J., Bryant, R. G., & McDonald, P. (2003). Mapping intertidal estuarine sediment grain size distributions through airborne remote sensing. *Remote Sensing of Environment*, 86, 480–490. [https://doi.org/10.1016/S0034-4257\(03\)00126-3](https://doi.org/10.1016/S0034-4257(03)00126-3).
- 545
- 546
- 547 Méléder, V., Launeau, P., Barillé, L., Combe, J. P., Carrère, V., Jesus, B., et al. (2010). Hyperspectral imaging for mapping microphytobenthos in coastal areas. In M. Maanan & M. Robin (Eds.), *Geomatic solutions for coastal environments* (pp. 71–139). New York: Nova Science Publishers Inc.
- 548
- 549
- 550 Small, C., Steckler, M., Seeber, L., Akhter, S. H., Goodbred, S., Jr., Mia, B., et al. (2009). Spectroscopy of sediments in the Ganges-Brahmaputra delta: Spectral effects of moisture, grain size and lithology. *Remote Sensing of Environment*, 113, 342–361. <https://doi.org/10.1016/j.rse.2008.10.009>.
- 551
- 552
- 553
- 554 Smith, G. M., Thomson, A. G., Moller, I., & Krompkamp, J. C. (2004). Using hyperspectral imaging for the assessment of mudflat surface stability. *Journal of Coastal Research*, 20(4), 1165–1175. West Palm Beach (Florida). ISSN 0749-0208.
- 555
- 556
- 557 Sogreah, (2009). Diagnostic, étude d'opportunité et de faisabilité pour l'aménagement de la rive Nord de l'Authie. *Rapport provisoire. Communauté de Communes Opale Sud*.
- 558
- 559
- 560 Tucker, C. J. (1979). Red and photographic infrared linear combinations for monitoring vegetation. *Remote Sensing of Environment*, 8, 127–150. [https://doi.org/10.1016/0034-4257\(79\)90013-0](https://doi.org/10.1016/0034-4257(79)90013-0).
- 561
- 562
- 563 Ubertini, M., Lefebvre, S., Gangnery, A., Grangeré, K., Gendre, R. L., & Orvain, F. (2012). Spatial variability of benthic-pelagic coupling in an estuary ecosystem: Consequences for microphytobenthos resuspension phenomenon. *PLoS ONE*, 7, e44155.
- 564
- 565
- 566 Verpoorter, C., Carrère, V., & Combe, J. P. (2014). Visible, near-infrared spectrometry for simultaneous assessment of geophysical sediment properties (water and grain size) using the Spectral Derivative–Modified Gaussian Model. *Journal of Geophysical Research : Earth Surface*, 119, 2098–2122. hal- 01130379.
- 567

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Chapter 45

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